

# A probabilistic approach to estimating timber harvest location

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## Abstract

Determining the harvest location of timber is crucial to enforcing international regulations designed to protect natural resources and to tackle illegal logging and associated trade in forest products. Stable Isotope Ratio Analysis (SIRA) can be used to verify claims of timber harvest location by matching levels of naturally-occurring stable isotopes within wood tissue to location-specific ratios predicted from reference data ('isoscapes'). However, overly simple models for predicting isoscapes have so far limited the confidence in derived estimates of timber provenance. In addition, most use cases have limited themselves to differentiating between a small number of pre-determined location options. Here, we present a new SIRA data analysis pipeline, designed to infer the harvest location of a focal tree out of a continuous, arbitrarily large area. We use Gaussian Processes to robustly estimate isoscapes from reference wood samples, and overlay with species distribution data to compute, for every pixel in the study area, the probability of it being the origin of the examined timber. This is the first time, to our knowledge, that this approach is applied to determining timber provenance, providing probabilistic results rather than a binary outcome. Additionally, we include an active learning tool to identify locations from which additional samples would maximize the improvement to model performance, allowing for optimisation of field efforts. We demonstrate our approach on a set of SIRA data from seven oak species in the USA as a proof of concept. Our method can determine the harvest location up to within 520 km from the true origin of the sample and outperforms the state-of-the-art approach. Incorporating species distribution data improves accuracy by up to 36%. The future-sampling locations proposed by our tool decrease the variance of resultant isoscapes by up to 86% more than sampling the same number of locations at random. The method we present here greatly advances the toolset available for verification of

47 timber harvest location, will empower authorities worldwide in enforcing anti-deforestation  
48 legislation and will help protect natural resources.

49 **Keywords:** stable isotopes, origin traceability, timber provenance, illegal logging, isoscapes,  
50 Gaussian Processes

## 51 **1 Introduction**

52 Unsustainable exploitation of natural resources is the largest driver of terrestrial biodiversity loss  
53 after land-use change (Díaz et al., 2019) and a major conservation challenge globally. To  
54 prevent a sixth mass extinction (Barnosky et al., 2011), nearly 200 nations have recently agreed  
55 on a new set of targets and goals under the Kunming-Montreal Global Biodiversity Framework.  
56 In particular, Target 5 of the agreement includes the objective to "ensure that the use,  
57 harvesting and trade of wild species is sustainable, safe and legal, preventing  
58 overexploitation" (2022 UN Biodiversity Conference, 2022). Meeting this ambitious target will  
59 require overcoming a key element of unsustainable use of natural resources: the illegal harvest  
60 of threatened tree species.

61 Legal frameworks have been established to combat illegal logging and trade in illegally harvested  
62 timber, such as the Convention on the International Trade in Endangered Species (CITES), the  
63 US Lacey Act (amended 2008), the UK Timber Regulation (2021), the EU Deforestation  
64 Regulation (EUDR; 2023) and the Australian Illegal Logging Prohibition Act (2012). The new  
65 policies place additional traceability and reporting requirements on companies trading in wood  
66 and agricultural products (Dormontt et al., 2015). For example, the EUDR requires operators  
67 to record and report the coordinates of production location (forest or farm), and enforcement  
68 officials will be expected to scrutinize those claims of harvest location. Despite the

69 comprehensive legislation already in place and the international commitments under current  
70 adoption, enforcement of such regulations remains a challenge. Illegally harvested timber is  
71 shipped under false declarations of origin or mixed into legal shipments, and methods for  
72 verifying geographical location have so far only been able to determine the correct location out  
73 of a few pre-determined options, mostly at country-level resolution (Watkinson et al., 2022a;  
74 Horacek et al., 2009; Muñoz-Redondo et al., 2021). This challenge is greatly intensified by the  
75 new EUDR legislation adopting precise geographical location (GPS point or polygon for plot of  
76 land) as a determinant of the legal status of timber.

## 77 **1.1 Stable Isotope Ratio Analysis to verify provenance**

78 Well-established scientific techniques enable measurement of the chemical, anatomical and  
79 genetic features of plants from a tissue sample (Deklerck, 2023), with ever increasing precision  
80 and availability. When compared against a robust physical reference collection, these attributes  
81 of the tissue can be used to (in-)validate declared species and origin claims, and support  
82 enforcement officials in their efforts to detect, for example, illegally harvested timber or fraud in  
83 supply chains.

84 Stable isotope ratio analysis (SIRA) is one of the most promising technologies in this context.  
85 Several elements within biological tissues (mainly Hydrogen, Oxygen, Carbon, Sulfur, Nitrogen)  
86 have multiple naturally-occurring stable isotopes, whose ratios vary predictably across space, in  
87 correlation with environmental conditions (West et al., 2010; Siegwolf et al., 2022; Gay et al.,  
88 2022; Pederzani and Britton, 2019). The heavy isotopes of these elements do not undergo  
89 radioactive decay, and their proportion can be readily detected by mass spectrometry (Boner  
90 et al., 2007). The isotopic composition of elements incorporated into the tissues of a plant is  
91 determined by soil properties, climate, metabolic fractionation and other biotic and abiotic

92 conditions characteristic of the species and the habitat in which the individual grows (Siegwolf  
93 et al., 2022; Camin et al., 2017; Horacek et al., 2009; van der Sleen et al., 2017; Gay et al.,  
94 2022). Hence, differences in stable isotope ratios among individuals correspond to the  
95 environment they grew in, and can be used to discriminate between plants from different  
96 geographic areas. SIRA has proven useful in determining risk of illegally harvested material in a  
97 wide variety of contexts, for example, forest products (Watkinson et al., 2020; Boner et al.,  
98 2007), wildlife trafficking (Bowen et al., 2005; Koehler et al., 2019; Wunder and Norris, 2008;  
99 Vander Zanden et al., 2015), ivory trade (Van der Merwe et al., 1990; Ziegler et al., 2016),  
100 agricultural products (Camin et al., 2016; Saadat et al., 2022), fish/seafood (Cusa et al., 2022;  
101 Silva et al., 2021; Kroetz et al., 2020), precious metals (Kirk et al., 2003), and natural and  
102 synthetic illegal drugs (Kurashima et al., 2004; Casale et al., 2005), but without the spatial  
103 discrimination required by the new timber legislation.

## 104 **1.2 Modelling approach**

105 Current modelling practices for the use of SIRA to verify harvest location of both legally and  
106 illegally harvested forest products require improvement. The use of SIRA is currently limited by  
107 the simplistic models used, as well as by the limited number of reference samples used as input  
108 data for such models. Reference sampling campaigns are costly and budgetary needs are often  
109 underestimated, with sampling locations often taking into account relative ease of sampling  
110 rather than areas that yield a gain in model prediction accuracy (Schmitz et al., 2019). There  
111 has been considerable development of isoscapes ("isotope landscapes"), given that stable  
112 isotopic variation is a continuous spatial variable in nature (West et al., 2010). These isoscapes  
113 are geospatial maps that show the isotopic variation of the material of interest (West et al.,  
114 2010). While the potential of isoscapes for determining forest product origins has long been

115 recognized, few rigorous methods exist to achieve this task. The existing methods use simple  
116 prediction strategies such as linear regression (Watkinson et al., 2020, 2022b), which do not  
117 fully leverage the information contained in isotope ratio data.

118 Gaussian Process (GP) regression, which is closely related to kriging in geostatistics literature,  
119 is a class of flexible regression models which use the values in sampled points to estimate the  
120 values in surrounding points (Li and Heap, 2008; Deklerck, 2023; Williams and Rasmussen,  
121 2006). A key advantage of GP regression is that it can quantify the uncertainty of its own  
122 predictions based on the inferred spatial covariance structure of the samples. The importance of  
123 quantifying the uncertainty of predictions is increasingly recognized in safety-critical (Jankowiak  
124 et al., 2020) and forensic (Su and Srihari, 2010; Swofford and Champod, 2022) machine  
125 learning applications. Additionally, GP regression facilitates inference of a sparsely sampled  
126 variable of interest from variables that are highly correlated with it but more densely sampled  
127 (Adhikary et al., 2017; Kanankege et al., 2018). In the context of plant harvest location  
128 estimation, this translates to inferring stable isotope ratios from atmospheric drivers (such as  
129 precipitation, temperature and water vapor pressure) known to influence the stable isotope  
130 signal in wood (Horacek et al., 2009; Siegwolf et al., 2022). This then provides a powerful tool  
131 for predicting the isotopic composition in areas that have not yet been sampled. However,  
132 previous work on timber isoscapes used GP regression primarily as a spatial interpolation  
133 technique without a probabilistic interpretation (Gori et al., 2018; Watkinson et al., 2022a).

134 Others used approximate GP models to derive variance estimates for origin determination in  
135 animals (Ma et al., 2020; St. John Glew et al., 2019).

136 Here, we develop GP-based probabilistic machine learning models to infer timber harvest  
137 location by directly modelling timber isoscapes from SIRA data, with the aid of atmospheric  
138 predictors and species distribution data. We show that probabilistic modeling greatly enhances

139 the utility of SIRA in estimating the geographical origin of timber, and, assisted by a reference  
140 dataset (Gasson et al., 2021), can be used to guide future sample collection by identifying  
141 sampling locations that will minimize prediction uncertainty.

## 142 **2 Materials and Methods**

### 143 **2.1 Data sets**

144 We use data from 87 trees of the genus *Quercus*, sampled across the contiguous United States,  
145 as described in (Watkinson et al., 2020). Stable isotope ratio measurements were done  
146 following the protocol described in (Boner et al., 2007). Each entry contained stable isotope  
147 ratio measurements of oxygen  $\delta^{18}\text{O}$  (ratio between  $^{18}\text{O}$  and  $^{16}\text{O}$ ), hydrogen  $\delta^2\text{H}$  (ratio between  
148  $^2\text{H}$  and  $^1\text{H}$ ), carbon  $\delta^{13}\text{C}$  (ratio between  $^{13}\text{C}$  and  $^{12}\text{C}$ ) and sulfur  $\delta^{34}\text{S}$  (ratio between  $^{34}\text{S}$  and  
149  $^{32}\text{S}$ ) as well as the GPS coordinates of the sampled tree. As stable isotope ratios are largely  
150 driven by environmental conditions such as precipitation, temperature, humidity and so on,  
151 publicly available datasets for these factors are used to aid the inference of isoscapes. We used  
152 the following atmospheric data:  $\delta^2\text{H}$  and  $\delta^{18}\text{O}$  isotopic composition of precipitation (Bowen and  
153 Revenaugh, 2003), water vapor (Borbas, 2015) (found to be associated with  $\delta^{13}\text{C}$  by  
154 (Watkinson et al., 2020)), reflected shortwave radiation (NEO, 2023) and precipitation  
155 (multi-satellite) (Huffman et al., 2020), both of which were found to be associated with  $\delta^{34}\text{S}$   
156 (Watkinson et al., 2020). For each of those data types, we used monthly means averaged over  
157 a number of years to minimize the impact of weather patterns in specific years (see (Watkinson  
158 et al., 2020) for precise year ranges).

159 To inform the priors (probability distributions representing the prior belief on possible tree  
160 locations) of the models we develop, we used species inventory data across the natural range of

161 each species within the United States (Wilson et al., 2013), downloaded from:  
162 <https://www.fs.usda.gov/rds/archive/Catalog/RDS-2013-0013> on 09/12/2022. This  
163 data is available as species-specific raster layers of tree abundance at 250m resolution. We then  
164 used the function `project()` of the R package *terra* (Hijmans, 2022) to bilinearly aggregate  
165 abundance data so that it matched the spatial resolution of other spatial data in the pipeline.

## 166 **2.2 Model architecture**

167 Figure 1 presents an overview of the data sets and components comprising our model and  
168 output. We use a rectangular grid to represent the study area. Grid points are placed every  
169 0.125 degree latitude ( $\approx 14$  km) and every 0.06 degree longitude ( $\approx 4.3$ -6.0 km), which allows  
170 us to approximate spatial probability distributions with high accuracy. For every isotope ratio  
171 (IR), we fit a GP regression model to the training data to obtain the posterior mean and  
172 variance of the isotope ratio for every point of the grid (see Appendix for the full detail on  
173 implementation).

174 Gaussian Processes are a class of flexible regression models, which enable the modeler to  
175 quantify the uncertainty about specific predictions. A GP regression model assumes that the  
176 responses (in our case, isotope ratios) at different locations are jointly normally distributed  
177 (Gaussian). The model is defined by three elements: (1) the mean, for which we use a  
178 constant, (2) the covariance function for which we use the Matern function (Williams and  
179 Rasmussen, 2006) with separate scaling parameters for latitude and longitude and (3) the noise  
180 parameter. The choice of mean and covariance functions reflects prior knowledge and modelling  
181 assumptions about the regression problem. The covariance function expresses the amount of  
182 information about unobserved locations contained in nearby observed values. The function  
183 parameters as well as the noise parameter are estimated by maximizing the likelihood of the



184 training data, in contrast to standard kriging approaches in geostatistics literature, which use  
185 approximate techniques based on summary statistics. We use GPyTorch (Gardner et al., 2018)  
186 to efficiently find the maximum likelihood parameter estimates.

187 The input to the GP consists of the coordinates and/or the climate variable values at the grid  
188 point. For a combination of observed stable isotope ratios  $(y_O, y_H, y_C, y_S)$  (meaning  $\delta^{18}\text{O}$ ,  $\delta^2\text{H}$ ,  
189  $\delta^{13}\text{C}$ ,  $\delta^{34}\text{S}$ ), we compute the likelihood of this observation at every point in the grid, using the  
190 four GP regression models estimated in the previous step. This likelihood is the product of  
191 likelihoods for each isotope ratio as we assume independence between isotopes. Given the prior  
192 and the likelihood, we compute the posterior probability of each grid point being the harvest  
193 location of the sample by multiplying the prior and the likelihood and normalizing so that the  
194 probabilities sum up to 1. For ease of interpretation, the output is a map with highest-posterior  
195 density (HPD) regions indicated for several probability levels (15%, 30%, 50%, 75%, 90%,  
196 95%).

197 To incorporate atmospheric data into the isoscape we use monthly averages of the atmospheric  
198 variables listed in Section 2.1. We use a linear covariance term to model the covariance  
199 component corresponding to the variation in the respective atmospheric variables. The linear  
200 covariance function models a linear relationship between the atmospheric variable and the  
201 response and is mathematically equivalent to Bayesian linear regression (Williams and  
202 Rasmussen, 2006). The overall covariance function is then the sum of the spatial and linear  
203 terms and can be seen in Appendix A.

204 We use the spatial density maps developed by (Wilson et al., 2013) to design two prior  
205 distributions for sample origin that account for the spatial distribution of oak species. The first,  
206 which we call the *density prior*, holds that the probability of a sample originating at a grid cell is  
207 proportional to the basal area (average amount of area occupied by tree stems per unit of

208 space) recorded at the grid cell. The second, which we call the *range prior*, assigns equal  
209 probability to every grid cell where above-zero basal area has been recorded. In addition, both  
210 priors allow for a small probability that a sample might occur outside its observed range - we set  
211 that probability to 0.01 and diffuse it uniformly over all grid points within the contiguous United  
212 States where the species does not occur according to (Wilson et al., 2013).

### 213 **2.3 Performance evaluation**

214 We perform 5-fold cross-validation on the data set and report the average values of all  
215 performance metrics over all data points. Samples with incomplete or ambiguous species  
216 information and samples collected in botanical gardens outside of their species' native range are  
217 excluded from the test sets, but not from the training sets, resulting in a total of 74 test  
218 samples across the 5 folds. We report performance of our models as well as our implementation  
219 of the approach by (Watkinson et al., 2020) averaged across the five cross-validation folds.  
220 Rigorously evaluating the performance of our models is a non-trivial task as each model  
221 produces a distribution over possible locations, rather than a single location. For this reason, we  
222 have defined several metrics to investigate different aspects of probabilistic harvest location  
223 prediction:

- 224 1. Predictive log-likelihood and log-posterior: We report the log-likelihood and the  
225 log-posterior of observing the sample at its true origin. Both of those measure how well  
226 the model fits the test data.
- 227 2. Mode distance: We report the great circle distance between the true location and the  
228 *mode* of the posterior distribution, i.e. the highest scored location according to the  
229 model. This metric measures the accuracy of the highest-scored locations, but it does not

230 account for the amount of uncertainty in model predictions.

231 3. Mean absolute error (MAE): To investigate how distant our predicted locations are from  
232 the truth, we report the expected distance between the true location and a location  
233 sampled randomly from the posterior distribution returned by our model

$$MAE = \int_{\mathbf{x} \in A} d(\mathbf{x}_t, \mathbf{x}) p(\mathbf{x} | \bar{y}, S) d\mathbf{x}$$

234 where  $d()$  is the great circle distance between the two points and  $p(\mathbf{x} | \bar{y}, S)$  is the  
235 posterior probability of  $\mathbf{x}$  being the harvest location. A perfect prediction would have the  
236 distance of 0. This metric will favour predictions concentrated around the true location  
237 over equally dispersed predictions concentrated elsewhere. It will also favour less dispersed  
238 predictions generally. For the method of Watkinson *et al.*, which only outputs a region of  
239 plausible locations, we assume a uniform distribution within the region highlighted by the  
240 model. In practice, isoscapes often predict similar isotope ratio values at distant locations,  
241 so even a statistically efficient method might yield a high MAE value.

242 4. Area scored higher than the true location (ASH): The behaviour of MAE is influenced by  
243 the shape of the posterior distribution, which favours unimodal over multimodal shapes.  
244 We report the total surface area corresponding to the points that the model considers  
245 more plausible than the true origin of the sample.

$$ASH = \int_{\mathbf{x} \in A} I[\text{score}(\mathbf{x} | \bar{y}, S) > \text{score}(\mathbf{x}_t | \bar{y}, S)] d\mathbf{x}$$

246 where  $I(\cdot)$  is the indicator function that yields 1 when the statement is true and 0  
247 otherwise. For all GP models, the score is the posterior probability of harvest location,

248 whereas for the method of Watkinson *et al.* we take the score to be the negative of the  
249 minimum value of the threshold that results in the true location being included in the  
250 highlighted region. In contrast to MAE, this metric is likely to give a low value to a  
251 posterior distribution that is concentrated in several small areas as long as one of those  
252 areas contains the true location. For example, if the true location could be a county in  
253 New York or a county in West Virginia, this would give a low ASH but high MAE as the  
254 two counties are far apart.

## 255 **2.4 Guiding future sampling efforts**

256 Field sample collections are time-consuming and expensive. We can optimize future field  
257 collections using informed prediction of where additional sample points are most needed for  
258 increasing origin estimation accuracy. The isoscape variance estimates provided by GPs can be  
259 used to guide future sampling efforts, which in turn will maximize the performance of the  
260 model. This approach is known as *active learning* in the machine learning literature. Here, we  
261 propose a strategy to minimize the error of our isoscape estimates by carefully choosing future  
262 sampling locations.

263 Early attempts at efficient active learning in GPs involved collecting samples at points with  
264 highest response variance or, equivalently, picking a set of points that maximizes the entropy of  
265 responses (Cressie, 2015). Unfortunately, this approach tends to recommend collecting samples  
266 on the boundaries of the study area, which is wasteful as the newly collected samples improve  
267 isoscapes in a smaller fraction of the study area than if they were placed away from the  
268 boundary. This motivated researchers to propose several criteria for optimizing  
269 sampling (Guestrin et al., 2005; Ramakrishnan et al., 2005). Here, we adopt an approach similar  
270 to that of (Guestrin et al., 2005) with a few modifications designed to address the large size of

271 our spatial grid, which renders their original method computationally intractable for our data set.  
 272 We seek to maximize the *average* reduction in predictive variance across our study area that can  
 273 be achieved by adding a sample to the training set. With  $S$  the set of sampled locations and  $G$   
 274 the set of grid points, we define the information gain (IG) associated with adding a new point  
 275  $(\mathbf{x}^*)$  to the training data set as

$$IG(\mathbf{x}^*) = \sum_{\mathbf{x} \in G} [(\log(\sigma^2(\mathbf{x}|S)) - \log(\sigma^2(\mathbf{x}|S \cup \{\mathbf{x}^*\})))] \quad (1)$$

276 where the predictive variances are computed using Equation A.4. The algorithm then picks the  
 277 point in the grid that yields the highest IG. Importantly, the predictive variances depend only on  
 278 the sampling locations and not on the sampled values, so it is possible to sequentially propose  
 279 multiple sampling points before collecting the samples. Our method sequentially proposes  
 280 sample collection points until a user-specified number of samples is reached. We assume that  
 281 samples can only be collected in locations where at least one of the species is present. Thus,  
 282 grid points that lie outside every species range are excluded from the procedure. Our active  
 283 learning strategy requires repeatedly computing a large number of predictive variances for  
 284 varying training sets. To reduce computation time, we randomly downsample our grid to 15000  
 285 points before running the analysis. In addition, we assume that the reduction in variance is  
 286 negligible for grid points situated more than 15 degrees away from the newly sampled point  $(\mathbf{x}^*)$   
 287 in longitude or more than 7.5 degrees in latitude.

## 288 **3 Results**

### 289 **3.1 Model accuracy and comparison**

290 The plausible location areas identified by our models consisted of points within an average  
291 distance of 520-870 kilometers from the true location of the oak tree samples, depending on  
292 model settings. Even with a relatively small training data set of 69 – 70 training samples  
293 (depending on the cross-validation fold), our model is able to exclude the vast majority of the  
294 study area from consideration as a possible source of the focal sample. All our models  
295 outperform the state-of-the-art method for determining timber harvest location (Watkinson  
296 et al., 2020) in most or all metrics. Table 1 shows performance metrics for all the models on  
297 the test data set.

298 Incorporating species distribution information improves prediction performance for every model  
299 and every metric examined except the log-likelihood, which is computed independently of the  
300 prior. Informative priors improve MAE by 16% to 35% and ASH by 15% to 57% with most  
301 improvement for the pure spatial model and least for the spatial+atmospheric model. The more  
302 informative density prior gives better accuracy than the range prior according to all the metrics.  
303 Predicted probability maps for a few test points are shown in Fig. 2 and 3.

304 The spatial-only GP model gives the closest location predictions to the true location of the tree  
305 samples, except when a flat prior is used. In general, the spatial-only model and the combined  
306 spatial+atmospheric model give similar results on all metrics and outperform the  
307 atmospheric-only model in almost all settings. Somewhat surprisingly, the combined model does  
308 not outperform the spatial-only model. This might be due to the relatively small dataset used  
309 here or the choice of atmospheric predictors, and remains to be tested as we continue to expand  
310 our reference databases. The predictions of atmospheric GP models appear qualitatively

311 different from those from the purely spatial GP, perhaps because atmospheric model predictions  
312 emphasize geographical areas with distinct climate patterns, such as Appalachia or the Gulf  
313 Coast. Unsurprisingly, the purely spatial GP identifies areas that are more spatially cohesive but  
314 do not share any obvious physical features.

### 315 **3.2 Guiding future sampling efforts**

316 We investigated the performance of our active learning strategy on the US oak data set. For  
317 the spatial-only model, we let our method propose 10 new sampling locations to add to the  
318 training data set in the first cross-validation fold and computed the predictive variances before  
319 and after including the proposed locations.

320 The resulting isoscape standard deviation maps are shown in Figure 4. Our active learning  
321 strategy proposes sampling locations in currently undersampled regions with high predictive  
322 variance and sampling in those areas results in a visible improvement. The highest decrease in  
323 predictive variance was observed for  $\delta^2\text{H}$  while the lowest decrease was observed for  $\delta^{18}\text{C}$ . Most  
324 of the chosen locations are close to, but not at the boundary of, the allowed sampling area.

325 To investigate the efficiency of our active learning procedure, we compared isoscape variances  
326 resulting from active learning with those resulting from adding the same number of points  
327 sampled randomly from the allowed sampling area. We generated 100 such variance maps and  
328 compared the average variance (across the allowed sampling area) of those maps with the maps  
329 in Fig. 4. Appendix B shows the average predictive variances as a function of the number of  
330 points added for both random and active learning sampling strategies. We see that our active  
331 learning strategy results in a substantially faster decrease in predictive variances. After adding  
332 10 samples, the reduction in variance achieved by our active learning method is between 64%  
333 ( $\delta^{13}\text{C}$ ) and 86% ( $\delta^{18}\text{O}$ ) greater than the average reduction achieved by the same number of

334 random samples.

## 335 **4 Discussion**

### 336 **4.1 Harvest location estimation**

337 To halt illegal logging, to enforce timber regulations and to protect biodiversity in forested  
338 landscapes, we need to be able to accurately estimate timber harvest location. Although several  
339 examples exist of applying SIRA for timber origin questions (Gori et al., 2018; Watkinson et al.,  
340 2020; Kagawa and Leavitt, 2010), these approaches do not take full advantage of (1)  
341 atmospheric and species distribution datasets available or (2) state-of-the-art probabilistic  
342 machine learning models. In addition, many SIRA use-cases limit themselves to a classification  
343 problem (country X versus country Y) compared to a continuous assignment problem (true  
344 harvest location). In response to growing evidence of fraud in supply chains, legislation  
345 increasingly requires operators to trace back to plot (for example the EU Deforestation  
346 Regulation). Consequently, determining the true harvest location will likely become increasingly  
347 important. In this work we present a new computational pipeline which aims at taking  
348 advantage of both (1) and (2) while predicting the harvest location as a continuous variable.  
349 The accuracy of our models depends on the specific modelling approach and the data sets used.  
350 Using prior information about species distribution results in a considerable increase in accuracy  
351 regardless of which model is used by all metrics considered. The impact of adding species  
352 distribution data appears to be greater for the spatial-only model than models that use  
353 atmospheric information. This could be due to climate patterns influencing both species  
354 distributions (habitat suitability) and the values of the atmospheric variables that we  
355 incorporated in our models, which renders species distribution information more redundant once



356 atmospheric variables have been included in the model.

357 Within timber tracing literature, our method bears the most resemblance to the work  
358 of (Watkinson et al., 2020), which uses linear regression to predict isoscapes based on  
359 atmospheric data. Their approach assumes a constant variance across the study area. In  
360 contrast, our method estimates the predictive variances based on the spatial covariance  
361 structure learned from the reference data, which enables us to translate differences in sampling  
362 density across regions into varying levels of confidence in isoscapes across space. Our method  
363 also assumes a linear relationship between atmospheric predictors and isoscapes, but our GP  
364 formulation implicitly integrates over plausible values of regression parameters, which should  
365 lead to more robust predictions compared to standard linear regression. In addition, our  
366 approach makes use of species distribution data, which yields substantially improved predictions  
367 compared to uninformative priors. Finally, our approach enables us to propose locations for  
368 further sample collection that maximize the utility of the samples.

369 The estimation of spatial covariance structure has recently attracted attention in animal stable  
370 isotope studies. (Ma et al., 2020) recently proposed a method that uses probabilistic  
371 precipitation isoscapes derived from a GP (Courtiol et al., 2019), which are then calibrated to  
372 produce isoscapes for the species of interest. (St. John Glew et al., 2019) introduced a model  
373 combining spatial and environmental effects using a novel likelihood approximation for isoscape  
374 estimation, though the main focus of their work is isoscape modelling, not origin estimation.

375 These approaches differ from ours in that 1) they rely on Laplace approximations for isoscape  
376 estimation rather than exact likelihood maximization; 2) they use ordinary least-squares  
377 regression to account for atmospheric predictors, whereas our method uses a Bayesian approach  
378 via a linear covariance term; and 3) they do not aim to actively improve isoscapes through  
379 additional sampling. A common feature between these models and ours is using a grid to

380 compute the posterior distribution of origins, which to the best of our knowledge was first  
381 considered by (Wunder, 2010).

382 Our current best performing model can estimate the harvest location for *Quercus* species to 520  
383 km across the east of the United States. Future field expeditions will lead to an improvement,  
384 especially if the identified priority locations are targeted (see 4.2). The presented model will be  
385 adapted to other use cases, with a focus largely on endangered tropical species which are under  
386 high logging pressure.

387 We expect that our models will be more accurate once more timber samples become available.

388 The size of the current data set of wood samples available to this study (87 samples) is quite  
389 small relative to the area of the contiguous United States, which inevitably results in large  
390 predictive variance in many areas. In addition to reducing uncertainty about undersampled  
391 areas, larger data sets (in the range of hundreds to thousands of samples collected from across  
392 the US) should also enable researchers to use more complex GP models, including models with  
393 heterogeneous noise (Binois et al., 2018), or deep GP models where the covariance function is  
394 modelled by a neural network (Wilson et al., 2016).

## 395 **4.2 Guiding future collection efforts**

396 Under the World Forest ID Programme (Gasson et al., 2021), tens of thousands of tree samples  
397 are being collected globally, and are being analysed by different techniques, including SIRA, to  
398 build georeferenced databases which can be used to identify timber harvest origin. Our active  
399 learning approach can be used to inform future sample collection efforts and increase model  
400 accuracy that can be achieved within a fixed sampling budget. This will be especially important  
401 in tropical regions, where reaching sampling sites can be difficult, time intensive and expensive.  
402 A good sampling design can substantially improve model performance (Contina et al., 2022),

403 and our method can be used to adapt sampling efforts as more data is analysed. Our current  
404 approach focuses on minimizing predictive variances without considering the impact of newly  
405 sampled points on model parameters. Extending our approach to *non-myopic* sampling (Krause  
406 and Guestrin, 2007), which considers the impact on model parameters, would constitute an  
407 interesting future research direction. Another avenue for improving our approach would be to  
408 augment our IG criterion to reflect the varying investment in collecting samples as a function of  
409 the time, logistics, and financial cost of reaching the desired sampling location.

## 410 **5 Conclusion**

411 The accurate estimation of geographic origin of globally traded wood products is a critical step  
412 in combating illegal logging and associated trade, by supporting authorities' ability to verify  
413 claims made by traders at any supply chain node. In this work we presented a novel analytical  
414 pipeline that brings together and incorporates multiple data types and algorithms. This  
415 methodology is able to accurately predict timber product origin and can be used to optimize  
416 future field sampling to further increase accuracy and precision. We hope that this work will  
417 inspire more efforts to expand reference collections of wood samples, and that governments and  
418 companies will more routinely use the technological tools at their disposal to have more  
419 oversight over their supply chains and promote a more sustainable use of natural resources.

## 420 **6 Conflict of interest statement**

421 The authors declare that they have no conflicts of interest.

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Table 1: Mean test set performance for all the models used in the study. Best values across all models are shown in bold and underlined whereas values that are not significantly different from the best values (Wilcoxon signed-rank test,  $p=0.05$ ) are shown in bold. The Spatial-only GP combined with the density prior gives the highest predictive log-likelihood and log-posterior and the lowest MAE and ASH values for all priors used. The Spatial-only model outperforms the other models when range or density priors are used, while the Atmospheric+Spatial model performs best in terms of MAE and ASH when flat priors are used. The inclusion of species distribution information decreases MAE and ASH values for all models used. All of our models outperform the earlier method of Watkinson *et al.* (Watkinson *et al.*, 2020) on most or all metrics.

model	prior	log L	mode distance (km)	MAE (km)	log-posterior	ASH (km <sup>2</sup> )
Spatial-only	flat	<b><u>-6.964</u></b>	<b>433</b>	809	-9.582	470000
Spatial-only	range	<b><u>-6.964</u></b>	<b>435</b>	600	-9.537	327000
Spatial-only	density	<b><u>-6.964</u></b>	<b>400</b>	<b>520</b>	<b><u>-9.059</u></b>	<b><u>203000</u></b>
Atmospheric-only	flat	-7.362	531	870	-9.972	576000
Atmospheric-only	range	-7.362	505	606	-9.797	450000
Atmospheric-only	density	-7.362	534	567	-9.428	311000
Atmospheric+Spatial	flat	<b><u>-7.149</u></b>	<b>408</b>	794	-9.518	382000
Atmospheric+Spatial	range	<b><u>-7.149</u></b>	<b>399</b>	627	-9.431	315000
Atmospheric+Spatial	density	<b><u>-7.149</u></b>	<b>463</b>	536	<b><u>-8.978</u></b>	<b><u>213000</u></b>
Watkinson <i>et al.</i>	NA	NA	886	859	NA	691000

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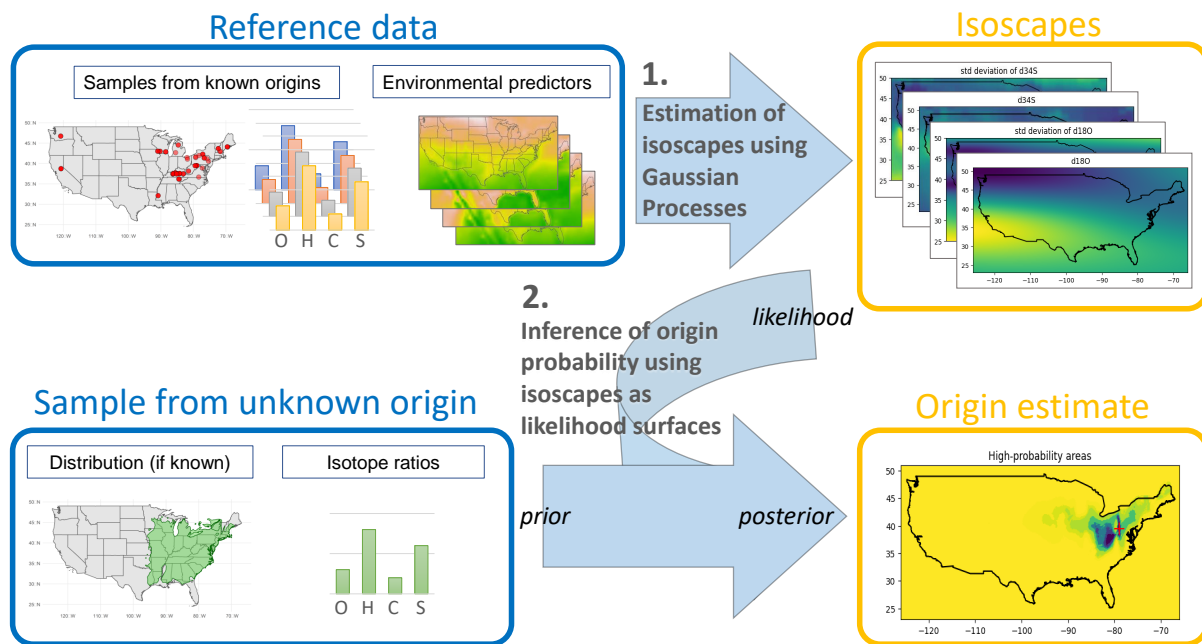


Figure 1: Model workflow. We use a training set of isotope ratios from trees collected at known locations and atmospheric data layers ("Reference data"). We fit a Gaussian process regression model to infer isoscapes and associated variance estimates, and compute the likelihood of observing the IR value for each element across the study area. To estimate the source of material with uncertain provenance ("Samples from unknown origin"), the isoscapes are then combined with prior information on the geographical distribution of the species, to yield a probability distribution of origin for the sample. We visualize predicted probability maps by plotting highest-posterior density regions for several probability levels (15%, 30%, 50%, 75%, 90% and 95%, dark blue to light green).

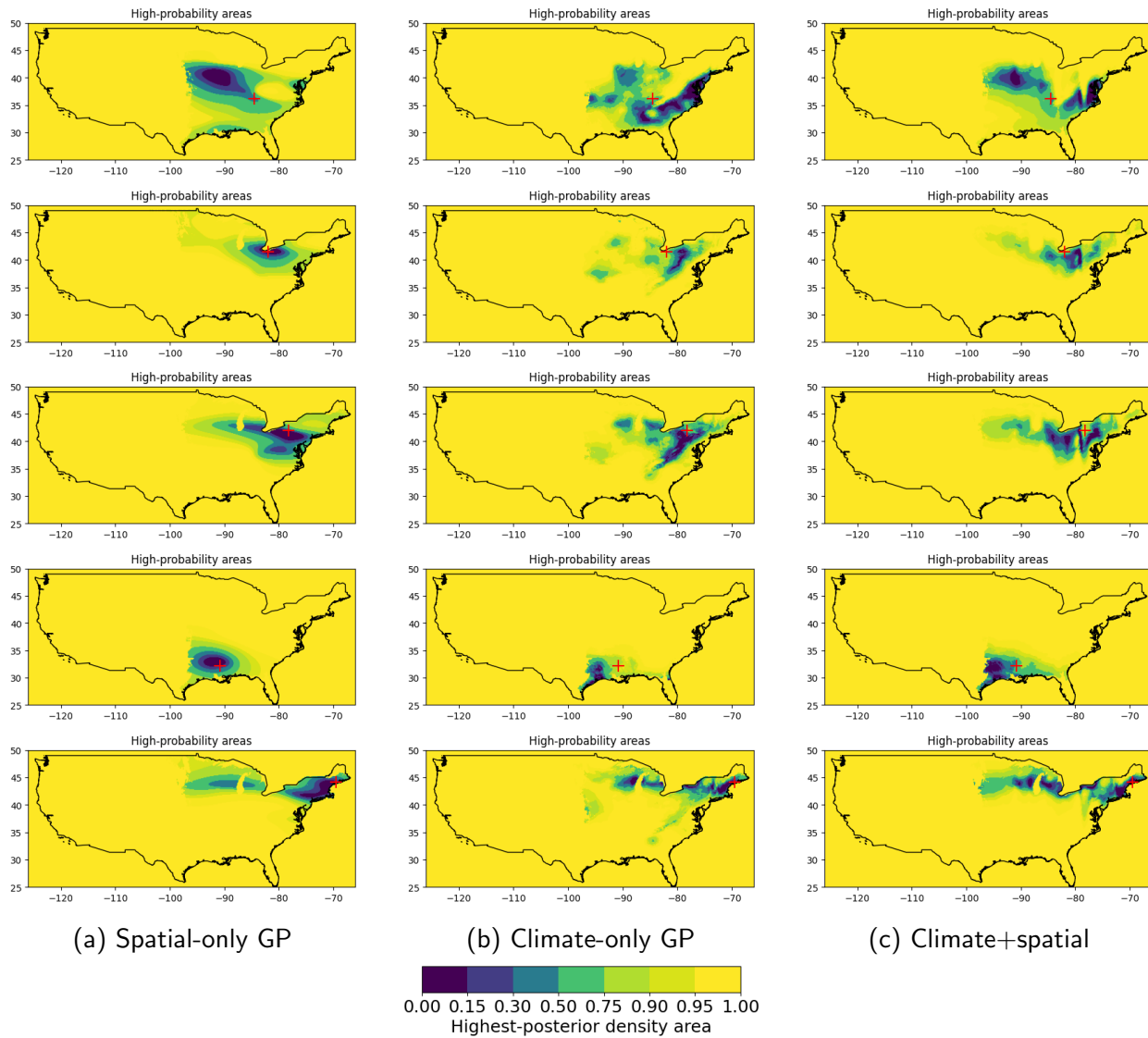


Figure 2: Harvest location predictions from the three models for 5 points from the test set using the range prior. Darker shades denote areas with higher posterior probability with thresholds set so that the total probability of the colored area is equal to a specified value (see color chart). The red cross indicates the actual location of the tree.

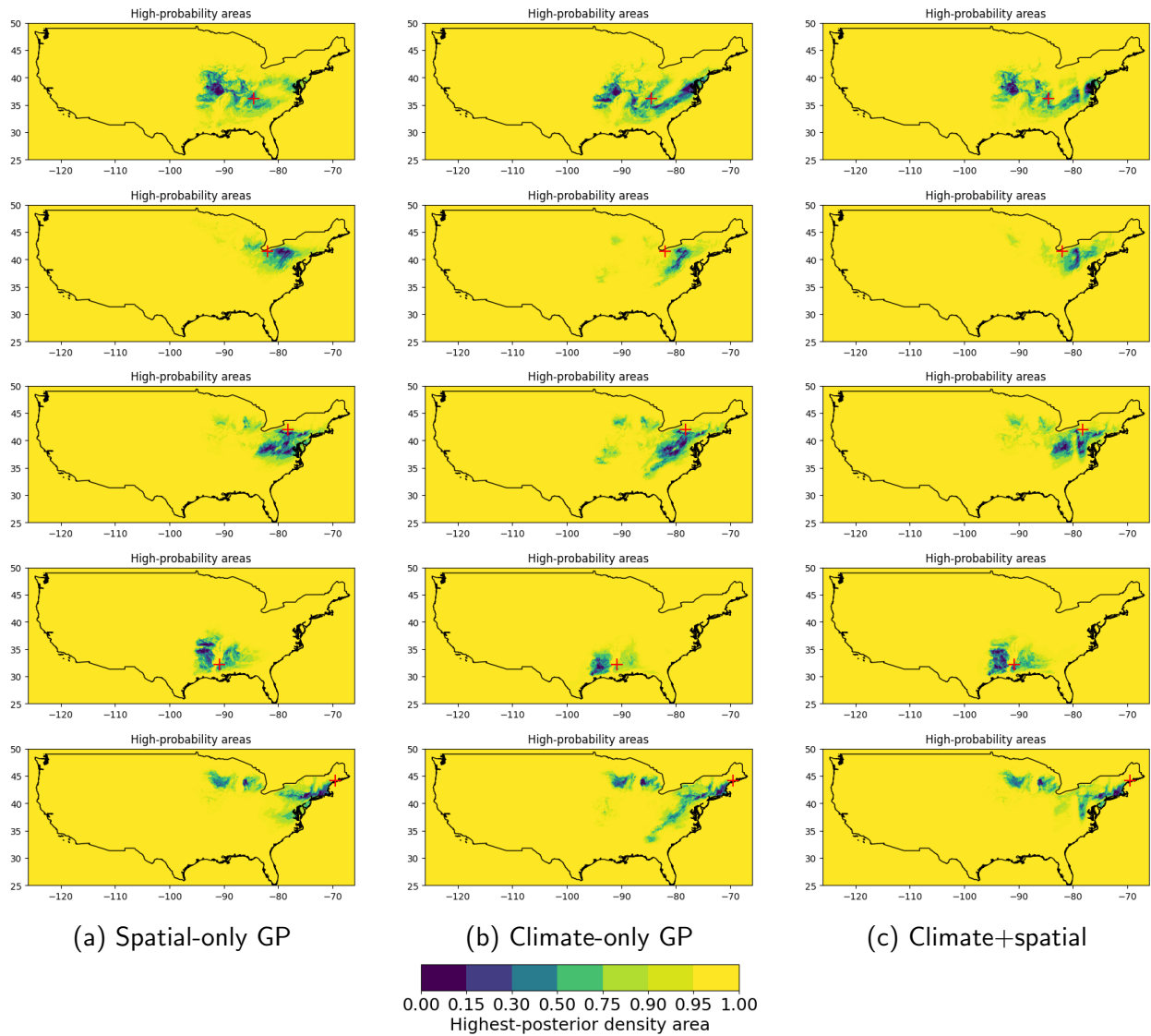


Figure 3: Harvest location predictions from the three models for 5 points from the test set using the density prior. Darker shades denote areas with higher posterior probability with thresholds set so that the total probability of the colored area is equal to a specified value (see color chart). The red cross indicates the actual location of the tree.

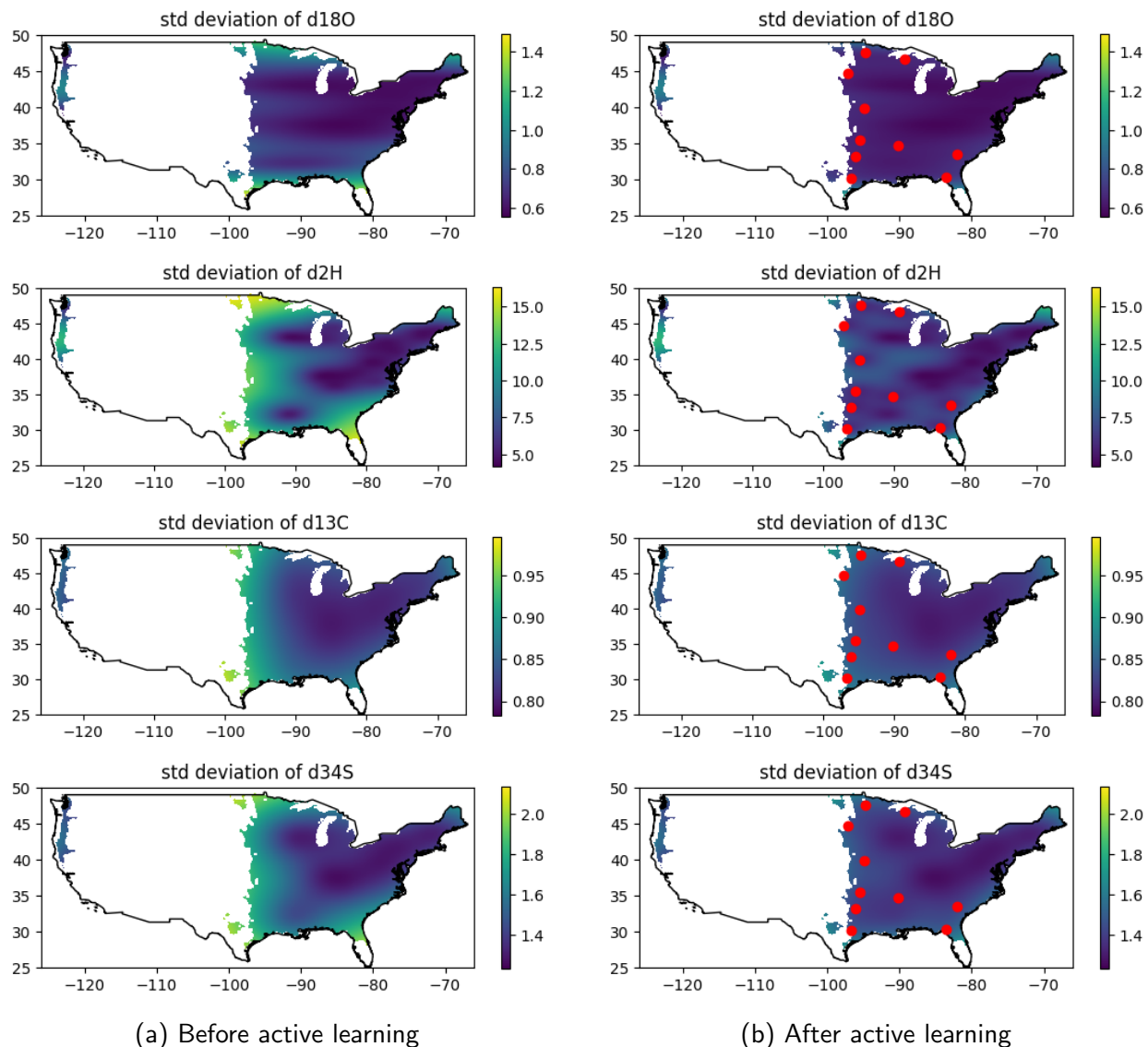


Figure 4: Maps showing predictive standard deviations for the four isotopes before and after adding 10 sample locations proposed by our active learning method for the spatial-only model. Standard deviations are only shown within the allowed sampling area, which is the union of ranges for the species in our data set. The red dots show the proposed locations. Our method proposes locations in areas with high predictive variance, particularly for  $\delta^2\text{H}$  and  $\delta^{34}\text{S}$ . Adding the proposed locations leads to a marked reduction of variance in the neighboring areas.



## 604 A More detail on Gaussian Processes

605 In the following, we give a brief overview of GPs. For an in-depth discussion, see Williams and  
 606 Rasmussen (Williams and Rasmussen, 2006).

607 GPs provide a flexible framework for regression, which enables the modeler to quantify the  
 608 uncertainty of specific inferences. A GP is a random process such that all of its marginals are  
 609 jointly normally distributed (Gaussian). Let  $\mathbf{x} = [x_{lon}, x_{lat}]$  be the GPS coordinates of a sample.

610 For any set of positions  $\mathbf{x}_1, \mathbf{x}_2, \dots, \mathbf{x}_n$ , the responses  $y_1, y_2, \dots, y_n$  at those positions are  
 611 assumed to be jointly normally distributed

$$\begin{bmatrix} y_1 \\ y_2 \\ \vdots \\ y_n \end{bmatrix} \sim \mathcal{N} \left( \begin{bmatrix} m(\mathbf{x}_1) \\ m(\mathbf{x}_2) \\ \vdots \\ m(\mathbf{x}_n) \end{bmatrix}, \begin{bmatrix} k(\mathbf{x}_1, \mathbf{x}_1) & k(\mathbf{x}_1, \mathbf{x}_2) & \dots & k(\mathbf{x}_1, \mathbf{x}_n) \\ k(\mathbf{x}_2, \mathbf{x}_1) & k(\mathbf{x}_2, \mathbf{x}_2) & \dots & k(\mathbf{x}_2, \mathbf{x}_n) \\ \vdots & \vdots & \ddots & \vdots \\ k(\mathbf{x}_n, \mathbf{x}_1) & k(\mathbf{x}_n, \mathbf{x}_2) & \dots & k(\mathbf{x}_n, \mathbf{x}_n) \end{bmatrix} + \sigma^2 \mathbf{I} \right) \quad (\text{A.2})$$

612 where:

- 613 1. The *mean function*  $m(\mathbf{x})$  describes the *a priori* expected value of the response  $y$  at  
 614 location  $\mathbf{x}$ . We use the constant mean  $m_c(\mathbf{x}) = \theta_m$  for all  $\mathbf{x}$ , where  $\theta_m$  is a parameter to  
 615 be estimated from the data.
- 616 2. The *covariance function*  $k(\mathbf{x}_1, \mathbf{x}_2)$  describes the *a priori* covariance between responses at  
 617 locations  $\mathbf{x}_1$  and  $\mathbf{x}_2$ . This is also a parameterized function. Popular choices of  $k$  are the  
 618 *squared exponential*  $k_{se}(\mathbf{x}_1, \mathbf{x}_2) = A \exp(-|\mathbf{x}_1 - \mathbf{x}_2|^2 / \rho^2)$ , or the Matern  
 619 function (Williams and Rasmussen, 2006), which both reflect the common assumption  
 620 that similar predictor values will lead to similar response values. In this work, we use the  
 621 Matern function with separate scaling parameters for latitude and longitude to model

622 spatial covariance.

623 3. The noise parameter  $\sigma^2$  models measurement error.

624 4.  $\mathbf{I}$  is the  $n \times n$  identity matrix.

625 We write  $\mathbf{y}$ ,  $\mathbf{m}$  and  $K$  to denote the responses, means and the covariance matrix of the training  
626 data, respectively, so that we can write Eq. A.2 as  $\mathbf{y} \sim \mathcal{N}(\mathbf{m}, K + \sigma^2\mathbf{I})$ . The choice of mean  
627 and covariance functions reflects prior knowledge and modelling assumptions about the  
628 regression problem. The function parameters as well as the noise parameter  $\sigma$  are estimated by  
629 maximizing the likelihood of the training data. We use GPyTorch (Gardner et al., 2018) to  
630 efficiently find the maximum likelihood parameter estimates.

631 After parameter estimation, the GP regression model can be used to predict responses at  
632 previously unseen data points. Let  $S$  be the locations and responses comprising the training  
633 data set. Since the responses at training and test points are assumed to be jointly Gaussian, the  
634 conditional distribution of the response at a test point  $\mathbf{x}^*$  given the training data is also  
635 Gaussian with mean

$$\mu(\mathbf{x}^*|S) = m(\mathbf{x}^*) + \mathbf{k}^*(K + \sigma^2\mathbf{I})^{-1}(\mathbf{y} - \mathbf{m}) \quad (\text{A.3})$$

636 where  $\mathbf{k}^* = [k(\mathbf{x}^*, \mathbf{x}_1), k(\mathbf{x}^*, \mathbf{x}_2), \dots, k(\mathbf{x}^*, \mathbf{x}_n)]$  is the vector of *a priori* covariances between  
637 responses at  $\mathbf{x}^*$  and training data points. The posterior variance of  $y^*$  is given by

$$\sigma^2(\mathbf{x}^*|S) = k(\mathbf{x}^*, \mathbf{x}^*) + \sigma^2 - \mathbf{k}^*(K + \sigma^2\mathbf{I})^{-1}\mathbf{k}^{*\top} \quad (\text{A.4})$$

638 - see (Williams and Rasmussen, 2006) for a derivation.

639 For a specific response value  $y^+$ , its likelihood of being observed at  $\mathbf{x}^*$  is just the Gaussian

640 probability density with mean  $\mu$  and variance  $\sigma^2$  found by applying Equations A.3 and A.4

$$p(y^* = y^+ | \mathbf{x}^*, S) = \frac{1}{\sqrt{2\pi\sigma^2(\mathbf{x}^*|S)}} \exp\left(\frac{-(y^+ - \mu(\mathbf{x}^*|S))^2}{2\sigma^2(\mathbf{x}^*|S)}\right) \quad (\text{A.5})$$

641 For a sample  $\bar{y} = (y_O, y_H, y_C, y_S)$  of observed isotope ratio values (meaning  $\delta^{18}\text{O}$ ,  $\delta^2\text{H}$ ,  $\delta^{13}\text{C}$ ,  
642  $\delta^{34}\text{S}$ ), the Bayes' theorem gives the posterior distribution of possible harvest locations:

$$p(\mathbf{x}|\bar{y}, S) = \frac{p(\mathbf{x}) \prod_{i \in \{O, H, C, S\}} p_i(y_i | \mathbf{x}, S)}{\int_{\mathbf{x} \in A} p(\mathbf{x}) \prod_{i \in \{O, H, C, S\}} p_i(y_i | \mathbf{x}, S) d\mathbf{x}} \quad (\text{A.6})$$

643 where the probabilities  $p_i$  are computed from the GP models for the respective isotopes using  
644 Equation A.5 and  $A$  is the study area. The integral in the denominator is computed by  
645 averaging the probabilities over the spatial grid.

## B Active learning performance

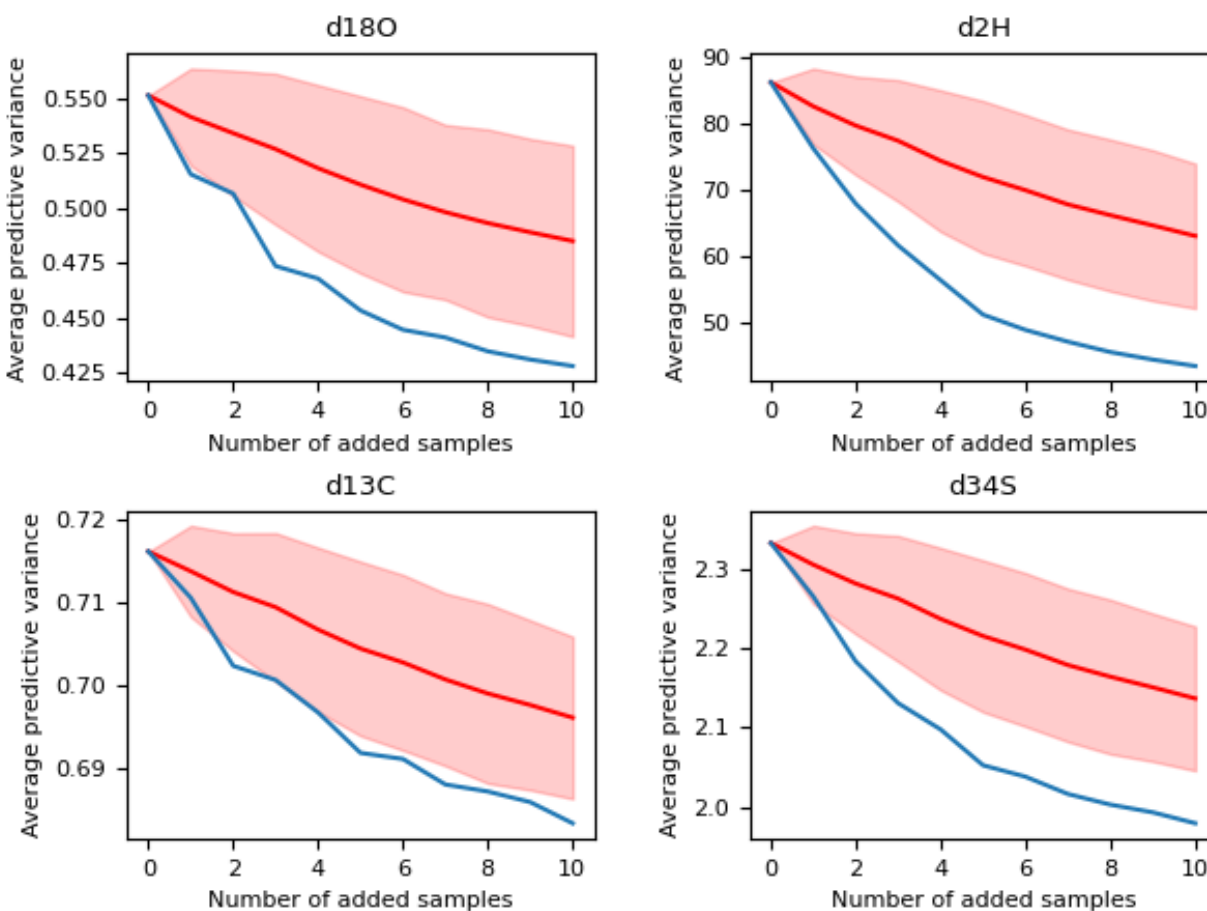


Figure 5: Average predictive variances for  $\delta^{18}O$ ,  $\delta^2H$ ,  $\delta^{13}C$  and  $\delta^{34}S$  as a function of the number of samples added to the base training data set; blue - active learning strategy; red - random sampling (shaded area denotes values within two standard deviations of the mean across  $n_r = 100$  simulations).